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Missing pieces in the annual cycle of fish ecology: a systematic
review of the timing of freshwater fish research focused on
juvenile Pacific salmon and trout

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20 **Abstract**

21 In recent decades, fish ecologists have become increasingly aware of the need for spatially
22 comprehensive sampling. However, a corresponding reflection on the temporal aspects of
23 research has been lacking. We quantified the seasonal timing and extent of freshwater fish
24 research reported in the literature. Since reviewing all prior work was not feasible, we considered
25 two different subsets. First, we compiled the last 30 years of ecological research on juvenile
26 Pacific salmon and trout (*Oncorhynchus* spp.) (n = 371 studies). In addition to the aggregate, we
27 compared groups classified by subject matter. Next, to characterize whether riverscape ecology
28 has embraced space at the expense of time, we compiled research across taxa for studies (n=46)
29 that sampled fish in a spatially continuous manner, i.e., those that followed Fausch et al.'s 2002
30 plea for a 'continuous view of the river...to understand processes interacting among scales.' We
31 found that the temporal distribution of ecological *Oncorhynchus* spp. research effort was biased
32 towards summer (40% occurred during June-August) and the month of June in particular, at the
33 expense of winter work (only 13% occurred during December-February). Riverscape studies
34 were also biased in temporal distribution toward summer (47% of studies) and against winter
35 (11%). It was less common for studies to encompass multiple seasons (43% of ecological
36 *Oncorhynchus* spp. studies and 54% of riverscape studies) and most were shorter than 4 months
37 (73% of ecological *Oncorhynchus* spp. studies and 81% of riverscape studies). These temporal
38 biases may cause researchers to overemphasize ecological phenomena observed during summer
39 and limit our ability to recognize seasonal interactions such as carry-over effects or
40 compensatory responses. Full year and winter studies likely hold valuable insights for
41 conservation and management.

42

43 **Introduction**

44 A key challenge in conservation is to understand how abiotic and biotic heterogeneity mediate
45 the function of ecosystems and the survival of biota that inhabit these environments. This
46 heterogeneity exists in both space and time, creating a shifting mosaic of physical and biological
47 conditions that has significant ramifications for biota [1]. Phenomena ranging from ontogenetic
48 niche shifts [2] to the stability of fisheries [3] can only be understood by jointly considering
49 interactions between space and time.

50 However, because resources are limited and characterizing stream heterogeneity is a non-
51 trivial task, it is often not feasible to study multiple dimensions of variation simultaneously.
52 Indeed, many fundamental concepts in stream ecology are either spatially or temporally focused.
53 For example, spatial patterns of biota are often described with minimal reference to time. This
54 applies to early work, such as the longitudinal zonation of fishes [4], but also the River
55 Continuum Concept [5] and the contemporary emphasis on spatially continuous sampling in
56 riverscape ecology [6]. Time is recognized as the “fourth dimension” of the riverscape [7] and
57 the intersections of various temporal and spatial scales has been noted as important [6].
58 However, in practice, the suffix “scape” often reflects a focus on spatial patterns with less regard
59 for temporal dynamics. Indeed, when Fausch et al. [6] called for stream ecology to embrace a
60 riverscape approach to achieve ‘a continuous view of the river’, the primary emphasis was on
61 closing gaps in space rather than time.

62 It is often recognized that short-term datasets can be inadequate because they fail to
63 capture historical levels of productivity (i.e. the shifting baseline) or reveal coarser scale
64 temporal patterning such as regime shifts [8]. Likewise, for cyclically patterned temporal
65 variation, interpretations may be misleading if they are based on a limited portion of a cycle. For

66 example, many fish switch between habitat types throughout the diel cycle [9] so only studying
67 animals during daytime may fail to capture important habitats. Similarly, refuge habitat
68 identified in summer may not represent refuge habitat for other seasons and stressors [10].
69 Riverine systems may exhibit extreme seasonal variation: water temperatures may range by more
70 than 20°C [11], and flows may vary by several orders of magnitude [12]. This variation strongly
71 affects not only fish and other aquatic organisms, but also the feasibility of field sampling. While
72 a temperature logger can effectively collect data every day of the year, the cost and logistical
73 challenges of sampling fish vary tremendously and can strongly govern when biological data are
74 collected. The intra-annual patterns of abiotic variables such as temperature and water quantity
75 are well quantified and describe the “regimes” of freshwater habitat. In contrast, biological
76 variables such as diet, growth, and survival exhibit analogous intra-annual variation, yet their
77 temporal patterning are quantified much less frequently. Extrapolating from data that pertain to
78 specific points in time can lead to misleading interpretations regarding how fish behave, what
79 locations or habitat types are important, and how much biomass ecosystems can produce [13,14].
80 This extrapolation of temporal data is particularly problematic in the study of mobile organisms
81 that undergo substantial physiological and ecological changes throughout their lifetimes [15].
82 The objective of this paper is to characterize the temporal attributes of fish ecology research to
83 elucidate potential data gaps that could guide future research.

84 Recent work on birds, amphibians, reptiles, and mammals found strong seasonal biases in
85 field research [16], but analogous work on fish has been lacking. The assertion that winter fish
86 ecology is an important, yet understudied portion of the research portfolio is not new [17];
87 however, this hypothesis remains unquantified. It was not feasible for us to screen the research
88 for all fish species during all life phases, so we limited our systematic review to a single genus of

89 fish: *Oncorhynchus*. We focused on juvenile Pacific salmon and trout in freshwater because they
90 are well-studied (providing us the power to detect trends in sampling), they live in highly
91 seasonal environments (which means an incomplete understanding of the annual cycle would be
92 a problem and is thus important to test for), and they are distributed across multiple continents
93 (thus representing a wide-spread species of interest). Here, we characterize the temporal aspects
94 of freshwater fish ecological research within the taxon of Pacific salmon and trout
95 (*Oncorhynchus* spp.) during the last 30 years. We characterized patterns in the seasonal timing
96 and extent of ecological field studies and considered how these patterns varied across three focal
97 topics: fish-habitat interactions, trophic ecology, and spatial distribution. Additionally, we
98 assessed whether spatially comprehensive sampling has come at the expense of time by
99 reviewing the timing of riverscape studies across all fish taxa.

100

101 **Materials and methods**

102 **Data screening**

103 To determine whether and to what extent temporal biases are present in fish field research, we
104 conducted a systematic review of two areas: 1) research within the *Oncorhynchus* species during
105 the juvenile life stage and 2) research across fish species within riverscape studies. We defined
106 riverscape fish studies as those employing spatially continuous (or nearly so) sampling at high
107 extents so that multi-scale patterns could be revealed [6]. These studies are in contrast to the
108 more typical method of using a relatively small number of points to represent large extents.
109 Thus, we did not use spatial extent as a variable in our analysis or a definition of riverscape
110 study, because high extent typically comes through high ‘magnification factor’ *sensu* Schneider
111 [18] that precludes robust study of spatial heterogeneity [6]. We focused on three temporal

112 aspects of research: 1) what months and seasons research occurs, 2) the duration of studies, and
113 3) whether research spanned multiple seasons.

114 To examine our first dataset of *Oncorhynchus* research, we reviewed 13 journals that
115 commonly publish research on fisheries ecology as opposed to human consumption of fish.
116 Using the Web of Science database (last searched 4 February 2021), we performed the following
117 search: TS=(salmon OR salmonids OR Oncorhynchus OR salmonine OR Salmonidae) AND
118 SO=(CANADIAN JOURNAL OF FISHERIES "AND" AQUATIC SCIENCES OR Ecology OR
119 Ecology of Freshwater Fish OR Ecosphere OR Ecosystems OR Environmental Biology of Fishes
120 OR Freshwater Biology OR Hydrobiologia OR North American Journal of Fisheries
121 Management OR Oecologia OR PLoS ONE OR Science OR Transactions of the American
122 Fisheries Society) Indexes=SCI-EXPANDED, SSCI, A&HCI, ESCI Timespan=1988-2017. The
123 past 30 years was chosen to characterize the current patterns of research and to make the review
124 more feasible due to limited access to journals during earlier periods of publishing. We screened
125 the articles and selected those that dealt with the ecology of juvenile *Oncorhynchus* species
126 during freshwater residence. The juvenile life stages of fry, parr, and smolt were all included.
127 We included both observational studies and experimental studies conducted in the natural
128 environment. We did not include studies that did not sample fish in the field, laboratory studies,
129 studies occurring in estuarine or marine environments, or studies that collected physical or
130 biological habitat data but did not actually sample fish. Our search terms identified 6,439
131 articles. After screening for non-eligible articles, 371 were included for analysis in this study (S1
132 Fig).

133 For the riverscape review, using the Web of Science database (last searched 23 October
134 2020), we performed the following search: TS=(riverscape OR spatially continuous OR

135 longitudinal distribution OR Fausch et al. 2002) AND TS=(fish OR fishes OR salmon) AND
136 TS=(stream OR river OR freshwater OR lake) Indexes=SCI-EXPANDED, SSCI, A&HCI, ESCI
137 Timespan=1988-2017. We then examined every article and selected those that dealt with
138 spatially continuous or high spatial extent (at or above segment scale [6]) sampling that included
139 fish data collection. Our search terms identified 411 articles. After screening for non-eligible
140 articles, 46 were included for analysis in this study (S2 Fig).

141

142 **Data analysis**

143 We classified each publication for both the ecological dataset and the riverscape dataset by the
144 temporal characteristics of data collection to examine the timing and duration of research
145 reported in literature. First, we read the Methods section of each article and screened and
146 recorded the presence/absence of data collection in each month and season. We defined seasons
147 meteorologically as aligned with the calendar months of June 1-August 31 for summer,
148 September 1-November 30 for autumn, December 1-February 28 for winter, and March 1-May
149 30 for spring. Presence and absence within a single month were used rather than breaking
150 months up by solstice and equinox dates. This was done because the dates of the solstice/equinox
151 change annually, and precise sample dates were often not reported beyond the month and year.
152 Studies may encompass more than one month, therefore the number of data points for these
153 analyses are greater than the number of studies included in the review. Second, we quantified the
154 frequency of the number of meteorological seasons (1-4) that were included in these studies to
155 analyze temporal extent and consideration of inter-seasonal interactions (i.e., carry-over effects).

156 To explore whether temporal aspects of sampling differed among research areas, we
157 classified each study into three focal areas: 1) fish-habitat interactions and the impact of habitat

158 units and types on juvenile salmonid biology or behavior, 2) trophic ecology including fish diet,
159 foraging, and food web structure, and 3) spatial distribution including movement and landscape-
160 scale distribution. Studies examining fish growth and survival were often presented by
161 researchers as a function of some aspect of one of the three focal areas identified and were
162 classified accordingly. The temporal distribution and extent of sampling effort was then
163 quantified both collectively and by research category. Each study was only classified into one of
164 the three focal areas based on the main objective of the study. Studies that did not fall into one of
165 these four main categories were classified as “Other” and included in overall analysis but not the
166 category-specific analyses.

167

168 **Statistical methods**

169 We used Pearson χ^2 -tests in R 4.0.2 to test the null hypothesis that data were randomly
170 distributed across all categories (i.e. sample month, or number of months sampled per year). If
171 no biases were present, we would expect similar distribution across the categories tested. While
172 the test is objective, we acknowledge that the interpretation is subjective due to the assumptions
173 that all months and seasons are equally important and present equal stresses, limitations, or
174 opportunities for growth, fitness, and survival for juvenile salmonids.

175 Our analyses presented two potential issues: multiple comparisons and non-
176 independence. We used a Bonferroni correction to adjust our threshold for statistical significance
177 (i.e., the alpha value) [19]. We performed 12 tests, so our adjusted alpha was 0.004 (i.e.,
178 0.05/12). To ensure that our interpretation of the Pearson χ^2 test was robust to non-independence
179 (from single studies contributing to multiple months or seasons), we used a Monte Carlo
180 approach. Specifically, we performed simulations in which (for the studies with multiple months

181 or seasons) we randomly selected only one month or season to be included in the analysis and
182 then recalculated the χ^2 statistic and p-value. We repeated this 10,000 times. This simulation
183 arrived at the same results as the standard Pearson's test (S4 Fig).

184 We acknowledge that seasonality varies with latitude, elevation, and position in
185 watershed, so the ecological conditions associated with a particular month or season may vary
186 among locations (and thus among the studies in our paper). Thus, the implications of the
187 temporal biases we observed may be somewhat context dependent. Future studies could address
188 this.

189

190 **Results**

191 **Monthly temporal distribution of studies**

192 At a monthly resolution across all ecological topics within juvenile *Oncorhynchus* spp. studies,
193 we found that the most frequently represented month was 3-6 times more common than the least
194 frequently represented month (Fig 1). December was the least represented month across all
195 topics, while the summer months of June, July, and August were most common among topics.
196 The month of June had a significantly higher proportion of studies than the month of December
197 at 14% and 3%, respectively.

198

199 **Fig 1. Temporal distribution of juvenile salmon ecology studies.** Left column: monthly
200 distribution (left to right: January to December) of sampling effort for juvenile Pacific salmon
201 and trout studies from 1988-2017 for (A) all studies ($X^2=289.58$, $p < 0.0001$, $n=1476$,
202 median=119.5), (B) habitat studies ($X^2=97.421$, $p < 0.0001$, $n=413$, median=28), (C) trophic
203 ecology studies ($X^2=78.131$, $p < 0.0001$, $n=244$, median=18), (D) spatial distribution studies

204 (X²=53.67, p < 0.0001, n=439, median=27). Right column: seasonal distribution of sampling
205 effort for juvenile Pacific salmon and trout studies from 1988-2017 for (E) all studies
206 (X²=243.39, p < 0.0001, n=1476, median=345.5), (F) habitat studies (X²=84.482, p < 0.0001,
207 n=413, median=83), (G) trophic ecology studies (X²=56.295, p < 0.0001, n=244, median=57.5),
208 (D) spatial distribution studies (X²=45.258, p < 0.0001, n=349, median=81). The number of
209 studies for each month or season was calculated using presence or absence of research during
210 that time frame. Dashed horizontal lines are data median. Studies may occupy more than one
211 month or season. Seasons were defined meteorologically, but as whole months. Summer is
212 defined as the months June, July, and August; Autumn is defined as the months September,
213 October, and November; Winter is defined as the months December, January, and February;
214 Spring is defined as the months March, April, and May.

215

216 **Seasonal temporal distribution of studies**

217 Across all ecological topics within juvenile *Oncorhynchus* spp. studies, we found that 39-44% of
218 studies occurred during summer while only 10-15% of studies occurred during winter (Fig 1).
219 There has been little change in the temporal distribution of research efforts with the proportion of
220 winter studies remaining lower than summer studies (Fig 2).

221

222 **Fig 2. Seasonal study distribution over time.** Change in the proportional temporal distribution
223 (seasonal timing) of all studies published from 1988-2017 in 5-year increments.

224

225 **Monthly temporal extent of studies**

226 At a monthly resolution across all ecological topics within juvenile *Oncorhynchus* spp. studies,
227 we found that most studies had limited temporal extent across the annual cycle, with 71-75% of
228 studies containing data from 4 months or less (Fig 3). Less than 2-8% of studies across all topics
229 encompassed data from all 12 months of the year.

230

231 **Fig 3. Temporal extent of juvenile salmon ecology studies.** Left column: frequency of the
232 number of months per calendar year (1-12) found in juvenile Pacific salmon and trout studies
233 from 1988-2017 for (A) all studies ($X^2=670.07$, $p < 0.0001$, $n=371$, median=5.1), (B) habitat
234 studies ($X^2=173.55$, $p < 0.0001$, $n=108$, median=4.6), (C) trophic ecology studies ($X^2=120.92$, p
235 < 0.0001 , $n=60$, median=8.3), (D) spatial distribution studies ($X^2=173.01$, $p < 0.0001$, $n=89$,
236 median=5.1). Right column: frequency of the number of seasons per calendar year (1-4) found in
237 juvenile Pacific salmon and trout studies from 1988-2017 for (E) all studies ($X^2=230.95$, $p <$
238 0.0001 , $n=371$, median=17.8), (F) habitat studies ($X^2=80.296$, $p < 0.0001$, $n=108$, median=16.7),
239 (G) trophic ecology studies ($X^2=19.6$, $p < 0.001$, $n=60$, median=20.8), (H) spatial distribution
240 studies ($X^2=72.573$, $p < 0.0001$, $n=89$, median=14.6). The extent or duration was calculated by
241 counting the total number of unique months (in a calendar year) that were included in each study
242 and categorizing them by season as defined above. Data median is marked with a dashed
243 horizontal line. Studies were only represented once at their greatest monthly extent and greatest
244 seasonal extent.

245

246 **Seasonal temporal extent of studies**

247 Across all ecological topics within juvenile *Oncorhynchus* spp. studies, we found that 48-63% of
248 studies occurred during a single season while only 6-10% of studies encompassed field sampling

249 from all four seasons (Fig 3). Only 43% of all studies collected data from multiple seasons and
250 73% of studies were shorter than 4 months. Again, there has been little change in the temporal
251 extent of research efforts with the proportion of single-season studies remaining significantly
252 higher than multi-season or year-round studies (Fig 4).

253

254 **Fig 4. Seasonal study extent over time.** Change in the proportional temporal extent (number of
255 seasons included) of all studies published from 1988-2017 in 5-year increments.

256

257 **Riverscape studies**

258 Analysis of riverscape studies across fish species revealed wider biases in temporal distribution
259 at monthly and seasonal scales. The most frequently represented month was 8x more common
260 than the least frequently represented month (Fig 5). January and February were the least
261 represented months, while June, July, August, and September were most common. Summer
262 encompassed 47% of all riverscape studies while only 11% of studies occurred during winter
263 (Fig 5).

264

265 **Fig 5. Distribution and extent of riverscape studies.** (A) Monthly distribution (left to right:
266 January to December) of sampling effort for spatially continuous “riverscape” studies involving
267 all fish species from 1988-2017 ($X^2=69.089$, $p < 0.0001$, $n=158$, median=8); (B) seasonal
268 distribution of sampling effort for riverscape studies ($X^2=54.152$, $p < 0.0001$, $n=158$,
269 median=33); (C) frequency of the number of months per calendar year (1-12) found in riverscape
270 studies ($X^2=97.038$, $p < 0.0001$, $n=46$, median=3.3); (D) frequency of the number of seasons per
271 calendar year (1-4) found in riverscape studies ($X^2=18.174$, $p < 0.001$, $n=46$, median=22.83).

272 The number of studies for each month or season was calculated using presence or absence of
273 research during that time frame. Dashed horizontal lines are data median. Studies may occupy
274 more than one month or season. Seasons were defined meteorologically, but as whole months.
275 Summer is defined as the months June, July, and August; Autumn is defined as the months
276 September, October, and November; Winter is defined as the months December, January, and
277 February; Spring is defined as the months March, April, and May.

278
279 Monthly temporal extent was limited within riverscape studies as well. Spatially
280 continuous studies were almost entirely conducted during a limited amount of time: 81%
281 contained data from 4 months or less and only 4% of studies encompassed data from a full 12
282 months out of the year (Fig 5). Seasonal extent for riverscape studies was the one metric that was
283 more representative than the ecological studies we examined: 46% of riverscape studies occurred
284 during a single season, 35% occurred over two seasons, 9% occurred over three seasons, and
285 11% occurred during all four seasons (Fig 5).

286 287 **Discussion**

288 In our review of 371 ecological juvenile *Oncorhynchus* spp. studies and 46 riverscape studies
289 from the last 30 years, we observed strong biases in seasonal timing (distribution) and temporal
290 extent. Within research topics where seasonality is particularly relevant, we observed the same
291 general pattern of temporal bias; the period of summer was overrepresented in the study of fish-
292 habitat interactions, trophic ecology, and spatial distribution. Below we discuss these temporal
293 patterns of data collection and consider their potential causes and consequences.

294

295 **Bias in temporal distribution of studies**

296 The most conspicuous pattern in the data was the lack of research during winter. For example,
297 the month of December had less than one-quarter as many studies as that of June. Winter studies
298 represented only 10-15% of total ecological research and 11% of riverscape studies. Winter may
299 be tempting to overlook because it is generally a period of low biological activity in freshwater
300 ecosystems. Winter is typically the coldest time of year, limiting the scope for growth and
301 activity in aquatic poikilotherms. Further, winter is the darkest time of year, limiting primary
302 productivity [20] and the foraging opportunity for visual predators [13]. Indeed, many stream-
303 dwelling fishes tend to allocate energy to fat stores in anticipation of winter [21], suggesting it is
304 generally a period of negative energy balance. However, decreased activity does not mean that
305 understanding winter ecology is not important. For example, if fish rely on summer and fall fat
306 stores to survive winter, then any food intake during winter could reduce the risk of depleting
307 energy reserves. Further, in some systems winter growth rates may actually exceed growth rates
308 during other times of the year [22,23]. Recent work has shown juvenile salmonids utilizing
309 different habitats in non-summer seasons [24] and acknowledges that our understanding of non-
310 summer habitat suitability is lacking [25], for a variety of taxa [26]. Identifying winter foraging
311 opportunities, trophic pathways, and habitat use could provide insights into how fish survive
312 during this time of year [27].

313 In many systems, winter survival is hypothesized to be a limiting factor, or bottleneck, to
314 freshwater population productivity of juvenile Pacific salmon and trout [28]. Therefore, reducing
315 winter mortality is often an objective of largescale restoration efforts [29]. While winter
316 mortality has been widely documented, the mechanisms behind winter mortality remain
317 unidentified or poorly corroborated in many of these fish populations [30]. In one study,

318 increased winter mortality in fish was linked to low late-season growth rates and high depletion
319 of energy reserves accumulated prior to winter [31], emphasizing the need for multi-season
320 research to explain productivity limitations. Further, the validity of winter as a bottleneck for
321 salmonid and non-salmonid species has not been widely supported, partially due to habitat-
322 specific characteristics of each system [17], thus supporting the idea that more research is needed
323 to understand winter fish ecology including habitat use, foraging ecology, winter mortality, and
324 productivity bottlenecks.

325 While the focus of winter ecology in juvenile salmonids is often on over-winter survival,
326 the impact of winter on freshwater fish populations goes beyond winter mortality. For fish in
327 lakes, daily and seasonal migrations exploit a portfolio of habitats throughout the year [32].
328 Additionally, the impact of climate change on the length of warm and cold seasons may have
329 unexpected consequences for fish. For example, one study found that shorter winters affected
330 perch populations by altering reproductive timing, egg size, and hatch rate, thus reducing overall
331 reproductive success of females [33]. The shift in season length could also cause a de-coupling
332 of nearshore and pelagic food webs in lakes [34].

333 The lack of winter research contrasted with an overrepresentation of summer studies. The
334 concern we raise is that relying on summer-biased data could pose problems for conservation and
335 management, specifically when data violate the assumptions for models. For example, species
336 distribution models (SDM) are increasingly used in climate change adaptation and rely on the
337 assumptions that a species occurs in all suitable habitats and that a species only occupies a
338 portion of that suitable habitat due to constraining factors such as competition or predation [35].
339 Developing such models from temporally biased data would be valid only if the focal species
340 were sedentary and their habitat use did not vary over time. However, it is rarely possible to

341 confirm that a species meets these criteria without having temporally representative data (i.e.,
342 you cannot dismiss the possibility of winter habitat shifts without data on winter habitat use).
343 Using data from a limited period of time can cause SDMs to erroneously dismiss critically
344 important habitat [36,15]. For example, one study demonstrated that SDMs based on seasonally
345 biased data failed to identify the habitats needed to support both hibernation and reproduction in
346 bats [37].

347 Defining climate refugia for fish based on summer-biased data [38] could similarly leave
348 out critical habitats if fish exhibit seasonal movements and require multiple habitat types to
349 complete the annual cycle [39,15]. Recent work focused on coldwater fishes showed how
350 physiological growth potential shifts across river basins through the annual cycle, peaking once
351 during summer in colder tributaries, and twice, during spring and autumn, in warmer
352 downstream areas [15]. Migratory life histories can exploit this spatio-temporal heterogeneity
353 through seasonal movement [34]. However, emerging frameworks for coldwater fish climate
354 adaptation do not place value on downstream growth habitats, because these approaches evaluate
355 habitats based on crowd-sourced data on fish occupancy, which is strongly biased toward
356 summer months [38]. This favors habitats that are optimal during summer at the expense of
357 habitats that function during other seasons. Simulations and empirical studies show that portions
358 of lakes or rivers that are thermally stressful and vacant of fish during summer may be critical for
359 growth during other seasons, fueling migratory life-histories that in turn provision fisheries [34,
360 15].

361 Seasonal bias and lack of year-round study may also limit our ability to apply
362 physiological models to fish conservation. For example, bioenergetics models for largemouth
363 bass were found to overestimate winter metabolic costs when the models had been empirically

364 validated using summer conditions [40]. Further, there is increasing evidence that animals can
365 adjust their physiological rates and capacities in response to intra-annual variation such as food
366 availability [41]. This can further introduce error into models that predict metabolism based on
367 data collected during specific times of the year [42,43].

368 Our current classification system for longitudinal fish zonation is largely based on
369 summer sampling [4]. While recent decades have seen an emphasis on more spatially
370 representative fish sampling [44] and a movement towards multiscale analysis of spatial
371 distributions [45], this work tends to not be temporally representative. For example, spatially
372 continuous “riverscape” sampling has been transformative for our understanding of salmonid
373 spatial distributions [6], yet our results confirm that virtually all of this work is conducted during
374 summer or early autumn [46,47]. Thus, a remaining challenge for riverscape ecology is to
375 achieve a ‘continuous view of the river’ [6] not only in space, but also across the annual cycle.

376 While longitudinal patterning is inherently relevant to lotic ecosystems (because they are
377 linear networks), fish may also exhibit pronounced spatial patterning in lateral, and vertical
378 dimensions [25]. In temperate regions of the Pacific salmon range, floodplains may only be
379 connected and wetted during winter, so summer-biased sampling may hinder our ability to
380 understand the significance of off-channel habitat use. Where summer and fall are the wet
381 seasons (e.g., much of coastal Alaska), use of off-channel habitats may vary seasonally and
382 require temporally extensive sampling to understand key dynamics. For example, the spatial
383 patterning of juvenile coho salmon on a stream floodplain shifted over time, tracking shifts in
384 water temperature [48] caused by fluctuating water levels. Use of temporary aquatic habitats by
385 fish may be disproportionately important when they are available at the right place and time;
386 however, research is lacking to capture this ephemeral aspect of fish ecology [49].

387 The distribution of juvenile salmonids among channel-unit scale habitat types [50] may
388 also vary among months and seasons. For example, one study found that juvenile coho primarily
389 occupied backwater pools in spring, main-channel pools in summer, and alcoves and beaver
390 ponds in winter [51]. Distribution of juvenile salmonids in sub-habitats (e.g. riffles, pools,
391 backchannel ponds) can also impact fish growth and fitness through energetic costs and benefits
392 [52]. While fine-detail studies of fish distribution help identify quality salmonid habitat, our
393 analysis demonstrates that this data implicitly favors summer habitat and devalues winter habitat.

394

395 **Bias in temporal extent of studies**

396 While a bias against winter studies is seen in temporal distribution, a bias against full annual
397 studies is seen in temporal extent. Ecological *Oncorhynchus* spp. studies examining all four
398 meteorological seasons represented only 6-10% of total research. While it is possible that a small
399 subset of winter studies could encompass the entirety of knowledge needed to fully understand
400 the winter ecology of fish, recent research has demonstrated that this is likely not the case as
401 novel and important insights continue to be found when understudied times of the year are
402 further explored [26]. Research is heavily skewed toward shorter, single season studies: 73% of
403 all studies capturing 4 months or less of data and 57% of studies focused on a single season in
404 isolation. Within riverscape studies, 81% of research occurred during 4 or fewer calendar
405 months. These patterns are similar to patterns observed in the temporal characteristics of
406 mammal, bird, reptile and amphibian research [16]. While there is increasing recognition of the
407 value of long-term study [53], this usually means having multiple years or decades of data
408 collection. Our review shows that there is also a lack of temporal extent in terms of the annual
409 cycle. Lacking extent at this timescale leads to two issues. First, we are likely to temporally

410 extrapolate and draw conclusions based on a subset of the year (as discussed above) and second,
411 we will often lack the ability to identify interactions between different time periods, or carry-
412 over effects [16].

413 Carry-over effects from one life stage or season can have significant impacts on fitness
414 and survival of individuals and populations in subsequent seasons or life stages [54]. As climate
415 change and increasing water demands make summer more stressful for salmon in regions such as
416 the western United States, there is a strong need to understand how conditions during spring and
417 fall mediate the effects of summer stress on freshwater rearing capacity. The ability of fish to
418 survive negative energy balance during harsh summer conditions should depend on their ability
419 to store energy in spring and rebuild energy stores in fall. For example, over-winter survival of
420 juvenile salmon is often positively associated with larger body size at the onset of autumn [55].
421 There is evidence that ephemeral food subsidy pulses, such as salmon eggs during the adult
422 spawning season, can positively influence juvenile salmon growth rate and energy density as
423 long as 6 months after this ephemeral resource pulse has disappeared [56]. Whether juvenile
424 salmonids grow large enough to consume eggs depends on their emergence timing and early
425 growth opportunities [57]. Thus, small increases in the growth of fry during spring may
426 determine whether marine subsidies benefit parr during fall, influencing overwinter survival and
427 the size of smolts the following spring, which relates to subsequent marine survival [58].

428 Sampling during multiple seasons is more likely to capture any carry-over effects that
429 span pre-pulse, pulse, and post-pulse. Food availability, along with temperature, strongly affect
430 fish growth rates with extreme variation in growth between seasons [22,59]. Quantifying fish
431 growth and food resources at multiple points in time are essential to avoid bias in assumptions
432 and to identify ephemeral trophic pathways that could be disproportionately important during

433 that season or in subsequent seasons. Additionally, consequences of increased stress during one
434 season can be observed in subsequent seasons through differences in fish growth, behavior, and
435 survival [60,61]. Compensatory responses, such as growth rate and survival after a period of
436 starvation, may also not be fully realized for many months [62,63]. The lack of full annual cycle
437 research on Pacific salmon has likely hindered our ability to recognize inter-seasonal carry-over
438 effects and compensatory responses, which may become increasingly important in the future.

439 A core concept in landscape ecology, also applicable to aquatic ecology, is that species
440 often rely on habitat complementation, meaning they utilize different patches of space containing
441 different resources at different times in order to complete their life cycles [64,10]. The use of
442 habitat by juvenile salmonids shifts 1) seasonally as river conditions such as temperature
443 gradually change [51], 2) momentarily as a balance of energetic costs and benefits [65], 3)
444 ontogenetically as resource needs change [2] and 4) ephemerally, such as during discrete events
445 like floods or drought [10]. Without full annual studies, the effects of these stressors on fish (e.g.
446 energetic costs, food availability, competition, predation) are poorly understood. Habitat
447 restoration may be more successful if information is available to allow for targeting of the
448 limiting life stage or limiting habitat in salmonid productivity [66]. It is well-established that the
449 challenges faced by stream-dwelling fishes in winter are vastly different [67]. To best protect the
450 habitat supporting juvenile salmon and trout, more effort is needed to understand the importance
451 of winter ecology.

452

453 **Considerations**

454 The seasonal bias of research could potentially be a product of two human limitations:
455 environmental challenges and allocation of scarce resources. First, the summer months generally

456 present the least challenging environmental conditions for human access to salmon-bearing
457 habitat, particularly in the Pacific Northwest where a significant amount of fish research takes
458 place: low stream flow, warm temperatures, and minimal precipitation. Sampling fish in the
459 winter months can be particularly challenging, as snow, ice, and high flow events limit safe
460 access for researchers and lead to fish exhibiting behaviors that make them difficult to capture
461 (e.g. winter concealment, nocturnality). Second, academic calendars create a seasonal bias
462 towards summer field work by their very structure, allowing time for field work while classes are
463 on break during summer. Field projects outside of academia also often follow a summer-
464 intensive field season program due to the availability of field technicians who are often college
465 students. Institutional hiring policies can further exaggerate these patterns. For example, at our
466 institution students cannot work > 20 hours per week during non-summer months, and it costs
467 ~30% more to hire seasonal assistants that are not students (due to the need for a temporary
468 hiring agency). This makes non-summer field work considerably more expensive. Thus, a
469 combination of environmental challenges, logistical hurdles, and institutional culture make field
470 work more likely to happen in summer.

471

472 **Conclusion**

473 In recent decades, stream ecology has strongly emphasized the need for more spatially
474 comprehensive sampling of fish [6]; however, temporally comprehensive sampling has not
475 received the same attention. Mapping the entire riverscape can reveal rich, multiscale patterns,
476 but efforts typically fail to reveal how these patterns shift over time. Fish may not occupy every
477 meter of space available to them, but they do live in every second of time. Furthermore,
478 phenomena such as floodplain dynamics [1], seasonal movement [68], portfolio effects [69],

479 resource waves [70], and thermoregulation [71] are driven by the interaction between spatial and
480 temporal variation. While not prescriptive, we propose a short list of research topics that may
481 benefit from year-round research: diet and bioenergetics, terrestrial-aquatic interactions, species
482 distribution and climate change vulnerability, use of ephemeral habitats, portfolio effects and the
483 importance of stable and ephemeral resources, and the impact of carry-over effects on fish
484 survival. We hope that our review encourages researchers to allocate more of their effort to
485 understudied portions of the year, which likely hold valuable insights for conservation.

486

487 **Acknowledgements**

488 We thank W. White for assistance with statistical analysis. We sincerely thank the reviewers
489 whose critical reading and helpful suggestions greatly improved this manuscript.

490

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687

688 **Supporting information**

689 **S1 Fig. PRISMA 2009 flow diagram for *Oncorhynchus* studies.**

690 (DOCX)

691

692 **S2 Fig. PRISMA 2009 flow diagram for riverscape studies.**

693 (DOCX)

694

695 **S3 Fig. PRISMA 2009 checklist.**

696 (DOCX)

697

698 **S4 Fig. Monte Carlo simulations of the Pearson χ^2 test repeated 10,000 times.**

699 (TIFF)

700 **S1 Table. Articles included in *Oncorhynchus* systematic review.**

701 (CSV)

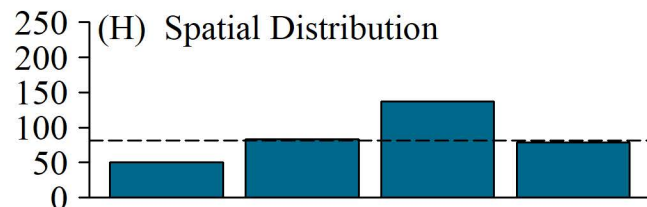
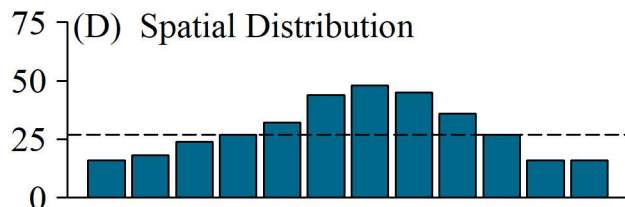
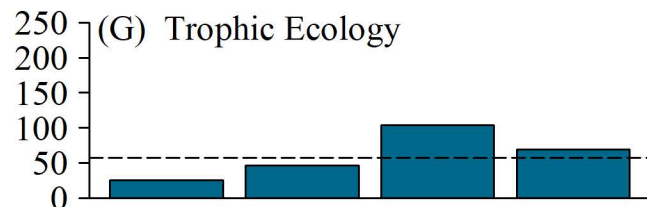
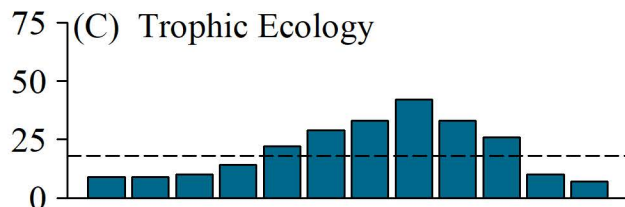
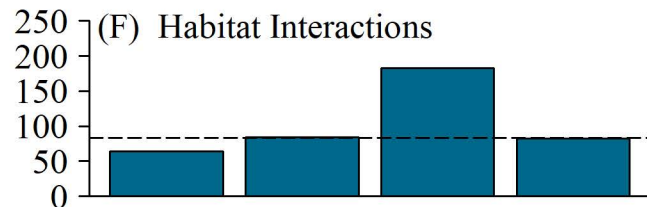
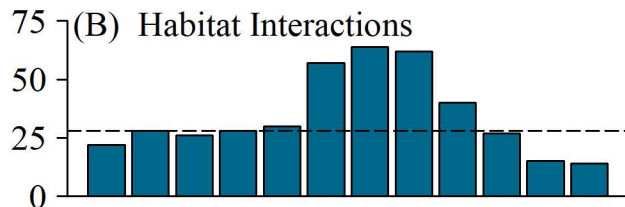
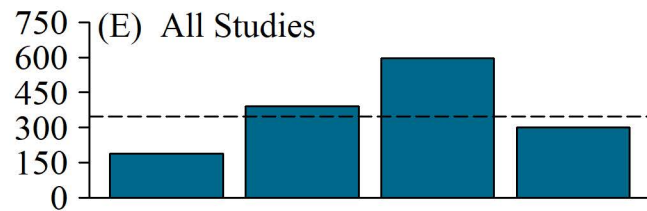
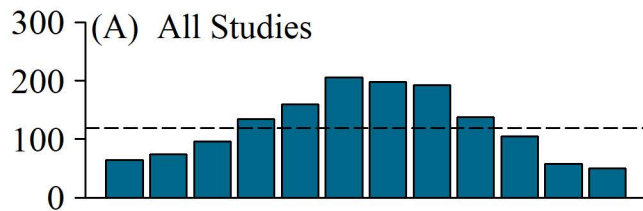
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703 **S2 Table. Articles included in riverscape systematic review.**

704 (CSV)

705

Number of Studies

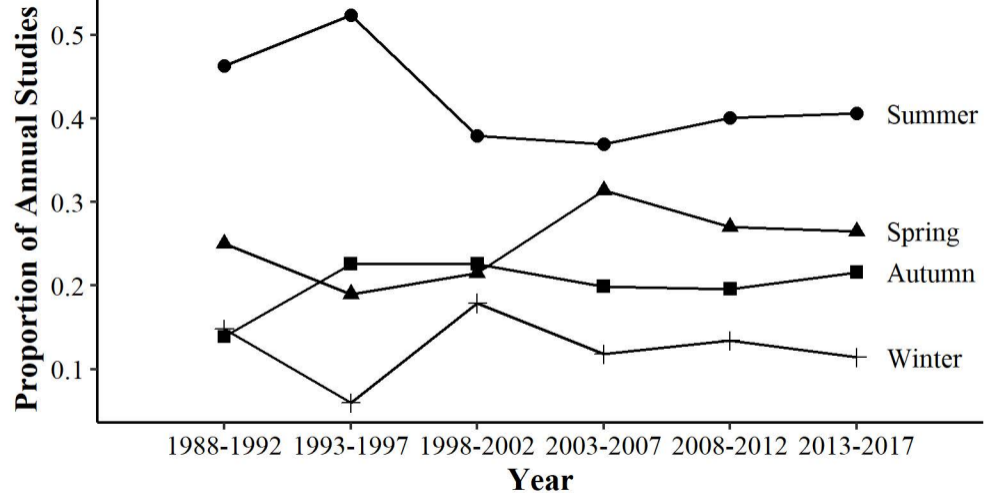


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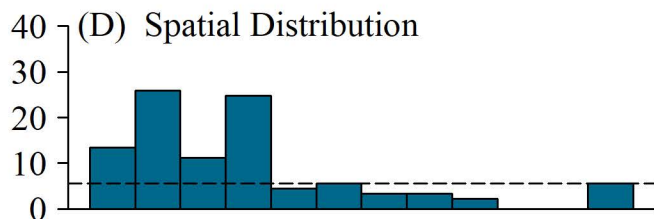
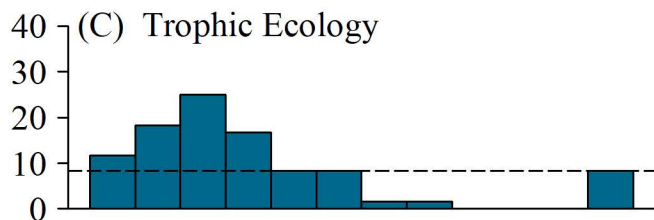
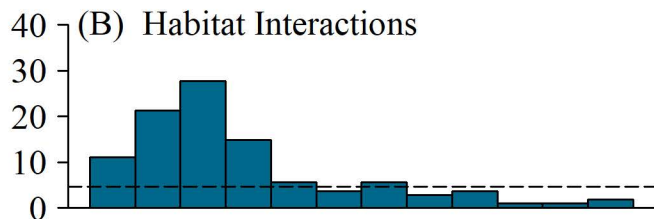
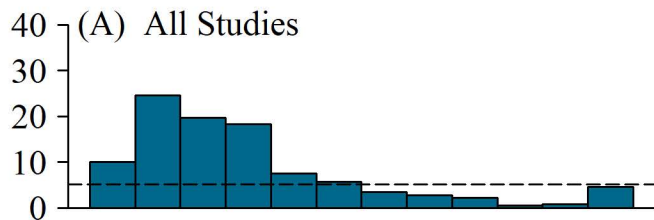
Wi Sp Su Au

Month

Season

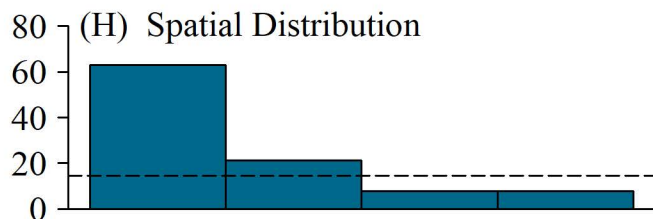
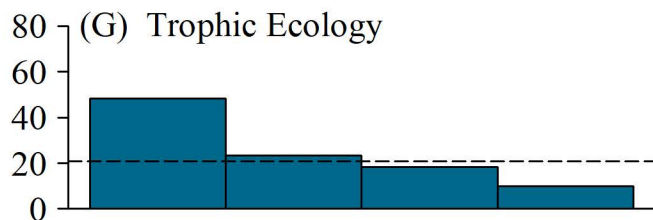
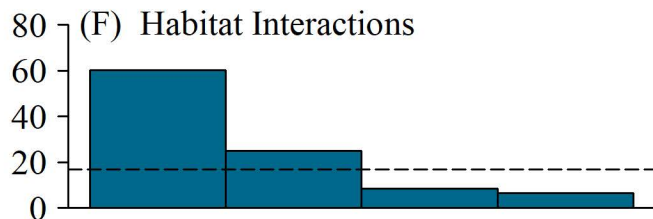
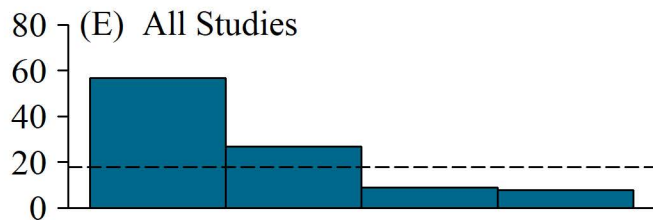


Percentage (%) of Studies



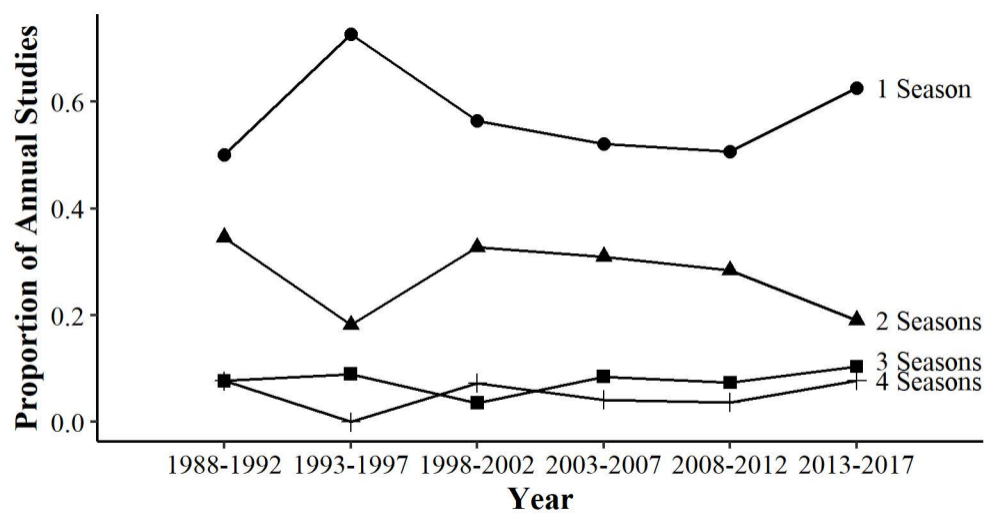
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Number of Months

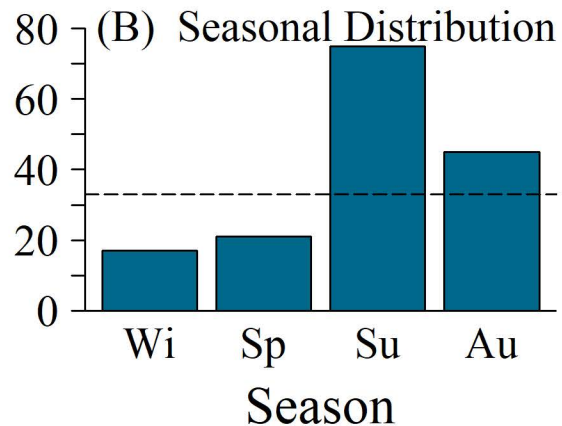
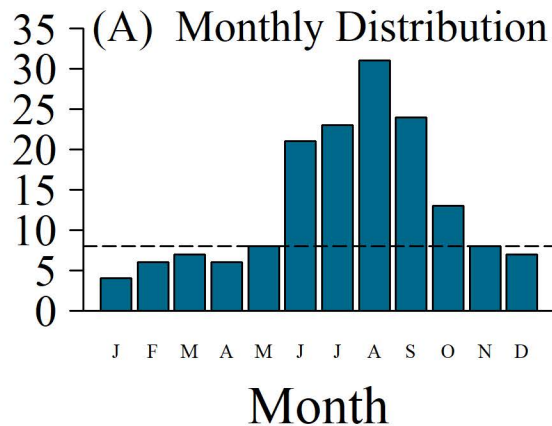


1 2 3 4

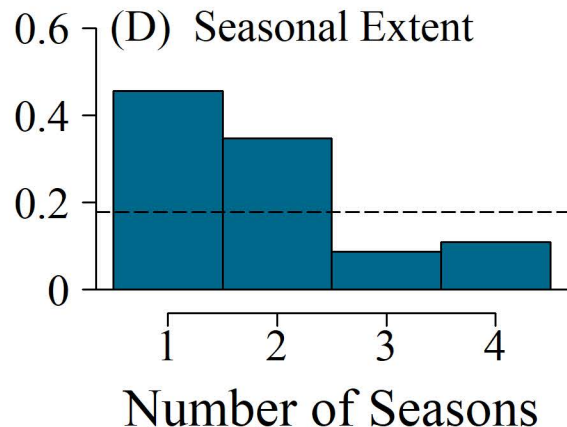
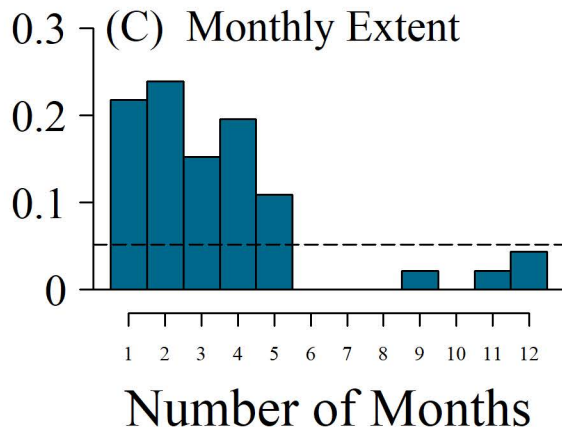
Number of Seasons



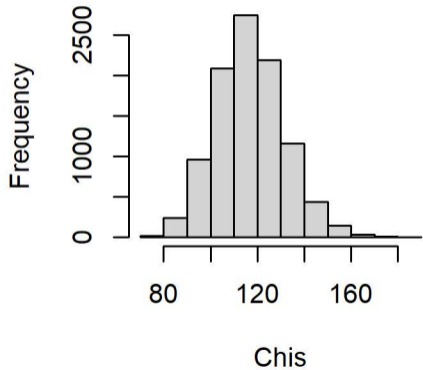
Number of Studies



Proportion of Studies



Histogram of Chis



Histogram of pvals

